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# Degradation of ciprofloxacin by UV and UV/H<sub>2</sub>O<sub>2</sub> via multiple-wavelength ultraviolet light-emitting diodes: Effectiveness, intermediates and antibacterial activity

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#### HIGHLIGHTS

- Ciprofloxacin is degraded by UV and UV/H<sub>2</sub>O<sub>2</sub> via multiple wavelengths UV-LEDs.
- 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> has the highest efficiency for ciprofloxacin degradation.
- UV irradiation only induces transformations of peripheral functional groups.
- UV/H<sub>2</sub>O<sub>2</sub> induces consecutive oxidation of core quinolone functional group.
- Incomplete degradation of ciprofloxacin reduced its antibacterial activity.

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#### G R A P H I C A L A B S T R A C T



## ABSTRACT

Although the efficiency and mechanism of the degradation of quinolone antibiotics by 254 nm UV radiation or UV/H<sub>2</sub>O<sub>2</sub> have been well elucidated, the same is not true for other UV wavelengths. The degradation of ciprofloxacin (CIP), a representative quinolone, was explored by UV and UV/H<sub>2</sub>O<sub>2</sub> using 255, 265, 280, 310 and 365 nm ultraviolet light-emitting diodes (UV-LEDs). The results of LC/MS<sup>2</sup> indicated that treatment at 280 nm UV/H<sub>2</sub>O<sub>2</sub> had the highest removal efficiency ([CIP] = 30  $\mu$ M, apparent rate constants reached 0.0759 min<sup>-1</sup>, half-time at 9.1 min) among these five wavelengths. Both the qualitative and quantitative analyses demonstrated that the intermediate abundance and distribution at 280 nm UV/H<sub>2</sub>O<sub>2</sub> were drastically altered in comparison to traditional 254 nm UV irradiation or photocatalysis. In the 280 nm UV-LED irradiation experiment, the primary intermediate was  $C_{17}H_{19}N_3O_4$  (m/z 330.1), which was generated by the substitution of the fluorine by a hydroxyl. In the  $280 \text{ nm UV-LED/H}_2O_2$ experiment, the dominant intermediate was  $C_{17}H_{18}FN_3O_4$  (m/z 348.1), in which a hydroxyl was added at the C=C bond of the cyclohexene moiety. The further oxidation of  $C_{17}H_{18}FN_3O_4$  was proposed to involve a consecutive oxidation pathway, following the order of CIP, C17H18FN3O4, C17H18FN3O5,  $C_{16}H_{18}FN_3O_4$  and  $C_{15}H_{18}FN_3O_3$ , which eventually destroyed the quinolone structure. Notably, the microbial analysis also proved that 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> degraded the moieties those are responsible for antibacterial activity. Based on these results, it was concluded that 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> can be used as a novel effective technology to improve the removal efficiency of quinolones in wastewater treatment. © 2016 Elsevier B.V. All rights reserved.

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## 1. Introduction

Quinolones are synthetic antibiotics that are widely used throughout the world as human and veterinary medicines. Quinolones are not biodegradable, and this feature results in their incomplete degradation in conventional biological wastewater treatment plants; therefore, residues of quinolones and their derivatives were discharged and detected in natural water bodies [1,2]. In recent years, many studies have confirmed that the existence of quinolones in the environment causes severe problems, including the induction of antibiotic resistance [3], the physiological teratogenesis of plants and algae [4], and the genotoxicity/carcinogenic potential [5]. In view of these facts, researchers are making efforts to develop cost-effective non-biological treatment methods to eliminate quinolones.

The popular advanced oxidation processes (AOPs), including Fenton oxidation [6], ozonation [7], peroxymonosulfate oxidation [8] and photocatalysis [9], were all shown to be effective for quinolones. Among these technologies, photolysis/photocatalysis using UV or visible light is very attractive because of its low cost, reduced use of chemicals, simple operation and wide application against a wide range of environmental pollutants. The decomposition of quinolones induced by UV photolysis was verified, but its mineralization rate was relatively low [10]. In contrast, photocatalysis using UV/visible light combined with catalysts generates hydroxyl radicals (OH'); thus, it has shown to lead to nonselective oxidation of quinolones, and its mineralization rate was higher [11–13].

Although multiple catalysts were developed and proven to be effective, the existing literature regarding the photolysis and photocatalysis of quinolones has never considered the optimization of the irradiation wavelength or the light source. Because quinolones are photosensitive molecules, screening for the optimal absorptive UV wavelength may improve its removal efficiency. Unfortunately, most studies have used a mercury arc lamp at a wavelength of 254 nm [14] or a xenon lamp with a continuous spectrum [15]. Moreover, these light sources have some drawbacks, such as high energy consumption, release of toxic heavy metals, frequent lamp replacement and fragility [16]. Thus, the exploration of more efficient UV wavelengths using eco-friendly light sources for the degradation of quinolones is urgently needed.

Within the last decade, the development of production techniques for solid-state devices has provided us with a compact, low-cost, low-energy, and environmentally friendly lightemitting diode (LED). The LED is free of toxicants and emits light at independent narrow bands (half bandwidth < 10 nm) [17]. An ultraviolet LED (UV-LED), which can emit UV light in the range of 200–400 nm, has been developed based on an aluminum gallium nitride material [18]. Thus, UV-LEDs can be used for the screening of the optimal wavelength for the degradation of particular organic contaminants. Recently, multifarious reactors based on UV-LEDs emitting different wavelengths were developed to degrade certain organic indicator compounds, including phenol [19], formaldehyde [20], methylene blue [21], methyl red [22] and even fulvic acids [23]. However, their ability to degrade antibiotics is still unknown.

In the present study, our research group designed a UV-LED module for the experimental degradation of antibiotics. Ciprofloxacin (CIP) was selected as an indicative quinolone, and the effectiveness of direct photolysis as well as  $UV/H_2O_2$  oxidation using UV-LEDs of different wavelengths was evaluated. The structures and generative pathways of the reaction intermediates were elucidated using a high-resolution tandem mass spectrometry. Finally, the reduction of antibacterial activity was evaluated by microbiological analysis. The findings of this study can be used to develop effective non-biological degradation methods for quinolones.

#### 2. Materials and methods

#### 2.1. Chemical reagents

Crystal CIP (98%, HPLC grade) and norfloxacin (98%, HPLC grade) were purchased from Sigma–Aldrich (USA) and stored at 4 °C. Analytical grade Na<sub>2</sub>SO<sub>3</sub> (98%) and H<sub>2</sub>O<sub>2</sub> (30%, v/v) were purchased from Sinopharm (China). HPLC grade acetonitrile and formic acid were purchased from Merck (Germany). All reagents were used as received. All of the solutions were prepared using ultrapure water (electrical resistivity: 18.2 MΩ) produced by a Milli-Q Advantage A10 system (Millipore, USA).

#### 2.2. UV-LED micro-module

A UV-LED irradiation micro-module was designed and assembled. This module consisted of five components, including a UV-LED array, power source, heat dissipation device, module framework and reactor vessel (Fig. 1). The UV-LED array was composed of a series of UV-LED chips, which had maximum emission peaks at 255, 265, 280, 310 or 365 nm, and the half-wave bandwidth of these chips was 10 nm (Fig. S1). The irradiating intensities of the UV-LEDs were measured using a HAAS-3000 light spectrum irradiation meter (Everfine, China). The average irradiation intensity of the 280 nm UV-LED chips was 0.023 mW cm<sup>-2</sup> at the surface of the reaction solution. The irradiation dose was calculated as:

$$\mathsf{Dose} = \mathsf{Int} \times T \tag{1}$$

where *Int* is the irradiation intensity and *T* is the irradiation time (s), and the irradiation dose has units of mJ cm<sup>-2</sup>. The irradiation intensities and doses of UV-LED chips of other wavelengths were also obtained using similar methods. In each array, a total of 16 UV-LED chips were set as a  $4 \times 4$  matrix upon the 6-cm diameter dish, and the dish was placed on the shaker. The arrangements were similar to all of the wavelength experiments. The module framework was designed by AutoCAD (Autodesk, USA) and was produced using the laser rapid prototyping technique. A customized circular quartz vessel (6-cm diameter) was used as the reactor vessel.

#### 2.3. UV-LED irradiation experiments

Twenty milliliters of CIP solution at  $10 \text{ mg L}^{-1}$  ( $30 \mu$ M) was spiked into the quartz vessel. In the UV/H<sub>2</sub>O<sub>2</sub> experiments, the initial concentration of H<sub>2</sub>O<sub>2</sub> was 10.2 mg L<sup>-1</sup> ( $300 \mu$ M). The pH value was maintained in the range of 6.5–7.2 using a 5.0 mM phosphate buffer solution. The experiments were performed in a customized incubator with constant temperature control. The solution was maintained at  $25 \pm 2$  °C, and uniformity was achieved by shaking the dish at  $60 \text{ r min}^{-1}$ . The reaction was initiated by turning on the UV-LED array. At a pre-defined time, Na<sub>2</sub>SO<sub>3</sub>, at a concentration that was stoichiometrically equivalent to the initial H<sub>2</sub>O<sub>2</sub> dose, was added to stop the reaction. Afterwards, 5 mL of the sample was transferred into a brown amber tube and then stored at 4 °C before sample analysis. The H<sub>2</sub>O<sub>2</sub>-only control experiments were included in the experimental design.

#### 2.4. HPLC separation

Before HPLC separation, all samples were mixed with isometric methanol that contained 1 mg L<sup>-1</sup> norfloxacin. The concentrations of CIP and the intermediates were verified by MS analysis using norfloxacin as an internal standard. The samples were injected into a LC-30AD liquid chromatograph system (Shimadzu, Japan) with a Waters Symmetry C-18 column ( $2.1 \times 150$  mm,  $3.5 \mu$ m) prior to the MS analysis. The injection volume was 10  $\mu$ L, and the mobile



Fig. 1. Micro-UV-LED module and its schematic diagram.

phase was a gradient elution of 0.1% formic acid water solution (mobile phase A) and acetonitrile (mobile phase B). The gradient elution was programmed as follows: 0–3.0 min, 10–90% B; 3.0–7.0 min, 90% B; 7.0–7.1 min, 90–10% B; 7.1–10.0 min, 10% B (40 °C, 0.3 mL min<sup>-1</sup>).

#### 2.5. Reaction intermediate analysis

The identification of intermediates was performed using a TripleTOF 5600+ high-resolution tandem mass spectrometry (HRMS) (Applied Biosystems SCIEX, USA) equipped with a Turbo V ESI ion source and a triple quadrupole time-of-flight (ToF) device. The instrumentation conditions are listed in Table 1. Nitrogen served both as the turbo and the collision gas. Mass calibrations and resolution adjustments on the quadrupoles and ToF were performed automatically using a  $10^{-5}$  M solution of polypropylene glycol (PPG) introduced via a model II Harvard infusion pump. The scan range was set at m/z 50–600. The data were analyzed using Peak-View and MasterView (Applied Biosystems SCIEX, USA). A systematic intermediate screening procedure is discussed in Supporting Information.

#### 2.6. Quantitative analysis of CIP and its intermediates

The quantitative analysis of CIP and its degrading intermediates was performed using a TripleQuad 5500 tandem mass spectrometry (Applied Biosystems SCIEX, USA). Most of the instrumentation conditions were similar to those of the 5600+ system, but the Collision-induced Dissociation (CID) energy was optimized for different intermediates. The scan mode was multi-reaction monitoring (MRM), and the monitoring ion pairs are listed in Table S1.

#### 2.7. Total organic carbon (TOC) and microbial analysis

To evaluate the mineralization efficiency, the TOC was measured using a VCSH-ASI TOC analyzer (Shimazu, Japan). The microbial experiment was performed to evaluate whether the reaction

Tá	able 1							
0	perational	parameters	of	tandem	mass	spectr	ometry	y.

Parameter (unit)	TripleTOF 5600+	TripleQuad 5500
Ion source mode	Positive	Positive
Scan mode	Full scan	MRM*
ESI needle voltage (V)	3500	3500
Turbo-gas temperature (°C)	350	350
Curtain gas pressure (psi)	40	40
Nebulizer gas pressure (psi)	35	35
Auxiliary gas pressure (psi)	40	40
Declustering potential (V)	60	60
CID energy (eV)	20 ± 15	-

MRM: multi reaction monitoring.

intermediates retained antibacterial activity. The agar plates were first inoculated with  $1.2\times10^8$  cfu mL $^{-1}$  *Escherichia coli* (ATCC 11303), and parallel 10-µL samples were separately spiked on agar plates. The plates were then incubated at 37 °C for 24 h, and the inhibition halo was measured. The diameter of the inhibition halo represented the antibacterial activity.

#### 2.8. Computational methods of CIP properties

The correlation of the molecular properties of CIP and its potential transformation pathway were computed by ChemBioDraw Ultra version 2015 [24]. Briefly, the molecular structure of CIP was drawn using ChemBioDraw and then copied to ChemBio3D to create a three-dimensional model. Subsequently, the structure was subjected to energy minimization by molecular mechanics until the root-mean-square gradient became smaller than 0.01 kcal mol<sup>-1</sup> Å. The bond energy and bond length of all atoms were calculated from the computed properties.

#### 3. Results and discussion

#### 3.1. Degradation kinetics of CIP

The result of the H<sub>2</sub>O<sub>2</sub>-only control experiment is presented in Fig. S2. A slight variation in the CIP concentration was observed, indicating that H<sub>2</sub>O<sub>2</sub> had a negligible effect on CIP. The removal effectiveness of the UV-LED-only irradiation experiments is presented in Fig. 2a, while the apparent rate constants and halftimes are summarized in Table 2. Illumination at 280 and 310 nm resulted in significant removal of the CIP, but the other wavelengths only induced slight variations. For  $10 \text{ mg L}^{-1}$ (30  $\mu$ M) CIP, the highest removal effectiveness reached  $\sim$ 66% after 60 min of UV-LED irradiation at 280 nm. This may be ascribed to the maximum absorption wavelength of CIP at 275 nm (Fig. S3). The apparent rate constant for the 280 nm UV-LED treatment reached 0.0181 min<sup>-1</sup>, with a half-time of 38.3 min. To the best of our knowledge, there has been no relevant research using different wavelengths of UV irradiation, except at 254 nm, for CIP degradation. In the previous study [14], low rate constants were observed for the direct degradation of CIP by a 254 nm mercury lamp, but no data regarding other wavelengths was reported.

The results of the UV/H<sub>2</sub>O<sub>2</sub> degradation experiments using UV-LEDs of different wavelength are presented in Fig. 2b. Obviously, treatment using 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> was the most effective, followed by 310 and 365 nm. The removal efficiencies at 280, 310 and 365 nm UV-LED/H<sub>2</sub>O<sub>2</sub> reached ~99%, ~97% and ~73%, respectively, within 60 min. As shown in Table 2, the apparent rate constants in the UV-LED/H<sub>2</sub>O<sub>2</sub> system increased from 0.0067 min<sup>-1</sup> at 255 nm to 0.0759 min<sup>-1</sup> at 280 nm, and then decreased to 0.0191 min<sup>-1</sup> at 365 nm. The half-time of CIP in the 280 nm/H<sub>2</sub>O<sub>2</sub> system was



**Fig. 2.** Removal efficiency of ciprofloxacin. (a) UV only, (b) UV/H<sub>2</sub>O<sub>2</sub>, (c) TOC. Experimental conditions: solution temperature  $25 \pm 2$  °C, pH 6.5–7.2, [ciprofloxacin]<sub>0</sub> = 30  $\mu$ M, [H<sub>2</sub>O<sub>2</sub>]<sub>0</sub> = 300  $\mu$ M (only in UV/H<sub>2</sub>O<sub>2</sub> experiment). The removal efficiencies of different wavelengths have been calculated based on irradiating dosages. All the experiments were carried out in triplicate with error bars representing the standard error of the mean.

#### Table 2

Apparent rate constants and half-lives of CIP in various wavelength UV-LEDs treatment.

Reaction	Confidence coefficient	Rate constant (min <sup>-1</sup> )	Half-time (min)
255 nm only	0.8658	0.0003	2310.5
265 nm only	0.9952	0.0019	770.1
280 nm only	0.9988	0.0181	38.3
310 nm only	0.9953	0.0129	53.7
365 nm only	0.9541	0.0030	231.1
255 nm/H <sub>2</sub> O <sub>2</sub>	0.9125	0.0067	103.4
265 nm/H <sub>2</sub> O <sub>2</sub>	0.9939	0.0106	65.4
280 nm/H <sub>2</sub> O <sub>2</sub>	0.9958	0.0759	9.1
310 nm/H <sub>2</sub> O <sub>2</sub>	0.9909	0.0549	12.6
365 nm/H2O2	0.9388	0.0191	36.5

9.1 min, suggesting that CIP can undergo fast degradation. Compared with the results of the UV-LED-only experiments, the degradation effectiveness of UV-LED/H<sub>2</sub>O<sub>2</sub> increased at the same wavelength, which may be ascribed to the promotion of OH oxidation.

The variation of TOC is presented in Fig. 2c. For a  $10 \text{ mg L}^{-1}$  $(30 \,\mu\text{M})$  CIP solution, the apparent TOC value was  $5.972\pm0.346\ mg\,L^{-1}$  at time zero. After a reaction time of 60 min, only 7% and 23% TOC were removed in the 280 nm UV-LED-only and the UV-LED/H<sub>2</sub>O<sub>2</sub> system, respectively. These low mineralization efficiencies can be attributed to two factors: (1) a lack of  $H_2O_2$  and (2) the low irradiation intensity (fluence rate) of the UV-LED chips. First, the [H<sub>2</sub>O<sub>2</sub>]:[CIP] was 10:1 in the current study, which was lower than the basic stoichiometric molar ratio for total degradation ([H<sub>2</sub>O<sub>2</sub>]:[CIP] = 47:1), resulting in an incomplete degradation of CIP. Second, the photo-quantum yield in a UV-based reaction is related to the irradiation intensity of light source; thus, the low irradiation intensity  $(0.023 \text{ mW cm}^{-2})$  of UV-LED resulted in a low generating efficiency of hydroxyl free radicals. This may also contribute to the low mineralization efficiency. In the research that used high-intensity irradiation treatment at 254 nm (up to  $0.43 \text{ mW cm}^{-2}$ ), the mineralization efficiency was significantly higher (>60%) [25].

#### 3.2. MS<sup>2</sup> analysis of CIP

According to the apparent mass-to-charge ratio equation (Eq. (2)),  $[CIP+H]^+$  had a m/z of 332.1405. An extract ion chromatogram of m/z = 332.1405 was obtained from the HRMS data, and thirty-seven distinct fragments were observed (Fig. S4). To avoid redundancy, the top 10 fragments in term of intensity were selected (Table 3). Nine compositions (included  $[CIP+H]^+$ ) were directly assigned to these fragments.

$$\frac{m}{z} = \frac{M + nH}{n} \tag{2}$$

where m is the apparent mass of the targeted molecular ion, z is the electric charge of targeted molecular ion, m/z represents the apparent mass-to-charge ratio, M is the exact mass of the targeted molecule, n is the number of electric charge, and H represents the exact mass of a proton.

The molecular structure of CIP is presented in Fig. S5. The central functional group is a guinolone moiety, and the peripheral moieties include a carboxyl group, a cyclopropyl group, a piperazinyl ring and fluorine. CID can induce cleavages of these moieties, and their losses may be single or multiple, resulting in different characteristic fragments. The fragment with an observed m/z = 314.1298 had the highest intensity (set as 100%) and was assigned as  $[C_{17}H_{17}FN_3O_2]^+$ , which had an approximate 18 Da  $(H_2O)$  difference compared to  $[CIP+H]^+$ . The fragment with the third highest intensity (69.2%) was  $[C_{16}H_{19}FN_3O]^+$ (m)z = 288.1497), which had an approximate 44 Da (CO<sub>2</sub>) difference compared to [CIP+H]<sup>+</sup>. These two fragments were also observed in the previous literature [26], and they reflected the existence of the terminal carboxyl group. Their high intensities suggested that the neutral losses of CO<sub>2</sub> and H<sub>2</sub>O were the dominant cleavages of CIP under the current HRMS conditions.

The fragment m/z = 245.1079 was assigned as  $[C_{13}H_{12}FN_3O]^+$ , and it may be the residual moiety that had previously lost both the carboxyl (MW 43.9892 Da) and the cyclopropyl (MW 43.0548 Da) groups. The fragment  $m/z = 231.0576 ([C_{13}H_{10}FNO_2]^+)$ may be formed through loss of the piperazine ring ( $C_4H_7N_2$ , MW 83.0609 Da) and the hydroxyl (MW 18.0100 Da). Furthermore, the fragment m/z = 294.1237 ( $[C_{17}H_{16}N_3O_2]^+$ ) may evolve from the cleavage of the C-F bond and the hydroxyl, while the fragment  $m/z = 268.1447 ([C_{16}H_{18}N_3O]^+)$  may evolve from the cleavage of the C—F bond and the carboxyl. Of note, the fragment m/z = 205.0773 $([C_{11}H_{10}FN_2O]^+)$  may include a cleavage at the cyclohexene in the quinolone moiety. The last fragment, m/z = 204.0156 $([C_{10}H_5FNO_3]^+)$ , was generated by the loss of the piperazine ring and the cyclopropyl group. These characteristic fragments can provide useful information for subsequent interpretation of the intermediates.

#### 3.3. Structure elucidation of intermediates in UV-LED-only experiment

According to the published studies, four sites in CIP, including the C—F bond, the C=C bond in the quinolone moiety, the piperazine ring and the cyclopropyl, may be attacked in the UV system and  $UV/H_2O_2$  system [25]. The computed results of the molecular

Table 3
MS/MS ions observed in the spectrum of [CIP+H] <sup>+</sup> .

m/z		Intensity	Error	Assigned molecular	MW difference with	Molecular structure	Characteristic
 Observed	Calculated	(%)	(Da)	formula	[CIP+H] <sup>+</sup>		molety
314.1298	314.1305	100	0.001	$[C_{17}H_{17}FN_{3}O_{2}]^{+}$ ([CIP+H-H <sub>2</sub> O] <sup>+</sup> )	18.0104	F C C C C C C C C C C C C C C C C C C C	Hydroxyl
332.1405	332.1410	77.4	0.001	[CIP+H] <sup>+</sup> C <sub>17</sub> H <sub>18</sub> FN <sub>3</sub> O <sub>3</sub>	0		-
288.1497	288.1512	69.2	0.001	[C <sub>16</sub> H <sub>19</sub> FN <sub>3</sub> O]⁺ ([CIP+H−CO <sub>2</sub> ]*)	43.9905		Carboxyl
245.1079	245.0964	35.9	0.011	[C <sub>13</sub> H <sub>12</sub> FN <sub>3</sub> O] <sup>+</sup>	87.0323		Carboxyl,
221.0576	221.0605	24.2	0.012	$([CIP +H-CO_2-C_3H_7]^*)$	101.0226	F HN N	cyclopropyl
231.0576	231.0695	24.3	0.012	[C <sub>13</sub> H <sub>10</sub> FNO <sub>2</sub> ] ([CIP +H-H <sub>2</sub> O-C <sub>4</sub> H <sub>7</sub> N <sub>2</sub> ] <sup>*</sup> )	101.0826		nyuroxyi, piperazine
268.1447	268.1450	13.1	0.000	[C <sub>16</sub> H <sub>18</sub> N <sub>3</sub> O] <sup>+</sup> ([CIP+H–CO <sub>2</sub> –HF] <sup>+</sup> )	63.9955		Fluorine, carboxyl
314.8798 294.1237	- 294.1242	9.8 7.9	_ 0.000	– [C <sub>17</sub> H <sub>16</sub> N <sub>3</sub> O <sub>2</sub> ] <sup>+</sup> ([CIP+H–H <sub>2</sub> O–HF] <sup>+</sup> )	17.2604 38.0165		– Fluorine, hydroxyl
						HN	(continued on next page)

Table 3 (continued)

m/z		Intensity	Error	Assigned molecular	MW difference with	Molecular structure	Characteristic
Observed	Calculated	(%)	(Da)	formula	[CIP+H] <sup>*</sup>		moiety
205.0773	205.0777	5.3	0.000	$[C_{11}H_{10}FN_{2}O]^{*}$ ([CIP+H-C <sub>6</sub> H <sub>9</sub> NO <sub>2</sub> ] <sup>*</sup> )	127.0629	F HN	Quinolone
204.0156	204.0097	5.3	0.001	$[C_{10}H_5FNO_3]^*$ ([CIP +H-C <sub>3</sub> H <sub>7</sub> -C <sub>4</sub> H <sub>7</sub> N <sub>2</sub> ] <sup>*</sup> )	128.1246	F O O O O O O O O O O O O O O O O O O O	Piperazine, cyclopropyl

properties (Fig. S6 and Table S2) were also in agreement with these previous studies and Scheme 2. Due to their low bond energies, the cleavage occurred in the N(11)–C(12) and C(1)–F(8) bonds more easily than in other C—C, N—C and C—O bonds. Among all of the C—C bonds, the one linking C(20) and C(21) in the cyclopropyl group had the lowest energy and was the easiest to break. The screening of potential intermediates from the HRMS data was based on the possible transformations of these sites. Eventually, three intermediates were confirmed in the 280 nm UV-LED-only experiment. They were identified as  $C_{15}H_{16}FN_3O_3$  (*m*/*z* 306.1),  $C_{17}H_{19}N_3O_4$  (*m*/*z* 330.1) and  $C_{17}H_{18}FN_3O_4$  (*m*/*z* 348.1) (Table 4 and Scheme 1). The observed fragments of CIP and its intermediates are summarized in Table S3.

Intermediate  $C_{15}H_{16}FN_3O_3$  (A) had a MW of 305.1176 Da, which was sustained by a 26.0156 Da difference with that of intact CIP. As reported by other studies [27], intermediate A was desethylene CIP, which may be formed through the cleavage of  $C_2H_2$  (MW 26.0156 Da) at the piperazinyl moiety. The MS<sup>2</sup> spectrum showed that the primary fragments were m/z 288.1155, 268.1089, 218.0487 and 190.0538 (Fig. S7). The observation of this intermediate implied that UV-LED irradiation induced the direct destruction of the piperazine ring.

Intermediate  $C_{17}H_{19}N_3O_4$  (B) had a MW of 329.1376 Da. This intermediate had a 1.9956 Da difference from CIP, which implied that its transformation involved the substitution of some peripheral moiety. The fragments of intermediate B included m/z312.1351, 286.1553, 243.1127, 229.0602, and so on (Fig. S8), which had formation patterns similar to the m/z 314.1298, 288.1497, 245.1079 and 231.0576 fragments, respectively, in the CIP MS<sup>2</sup> spectrum (Table S3), indicating the presences of carboxyl, cyclopropyl, and piperazinyl groups. In the CIP MS<sup>2</sup> spectrum (Fig. S8), the fragments m/z 268.1447 and 294.1237 both represented the C–F bond. However, the corresponding characteristic fragments were absent in the MS<sup>2</sup> spectrum of intermediate B, suggesting that it may have lost the C—F bond. Therefore, the transformation from CIP to intermediate B may involve the elimination of a fluorine atom (MW 18.9984 Da) and the addition of a hydroxyl (MW 17.0027 Da) (Scheme 1). In the previous study using UV irradiation at 254 nm, an intermediate with a similar structure was also observed [14]. These results suggested that direct 280 nm UV-LED irradiation attacked the C–F bond, followed by the addition of a hydroxyl group on the quinolone.

Intermediate  $C_{17}H_{18}FN_{3}O_{4}$  (C) had a MW of 347.1281 Da. This intermediate had a 15.9951 Da difference from CIP, suggesting the addition of an oxygen atom or the substitution of a hydrogen

atom by a hydroxyl. The primary fragments had m/z 330.1235, 304.1443, and 261.1003 (Fig. S9), indicating the presence of both carboxyl and cyclopropyl groups (Table S3). Furthermore, the minor fragments m/z 247.0517 and m/z 310.1182 implied the existences of the piperazinyl ring and fluorine. These results suggested that the 15.9951 Da transformation may be located at the quinolone moiety. Therefore, intermediate C may have a hydroxyl substitution at Site 10 of CIP [28].

The relative intensity variations and abundance variations of these intermediates are presented in Fig. 3. The intensity of intermediate B increased to approximate  $2.0 \times 10^5$  at 50 min and then showed a decreasing tendency. Intermediate A increased to  $1.0 \times 10^5$  at 17 min and disappeared at 40 min. The variation tendency of intermediate C was relative inconsistent with a maximal intensity of  $3.2 \times 10^4$  at 10 min. Although the exact concentrations of these intermediates were not determined accurately, as we did not have authentic standards for them, these data are still important in evaluating the variation tendency and abundance of these intermediates. These results suggested that UV-LED irradiation mainly induced transformations of the peripheral moieties. Undoubtedly, the hydroxyl substitution of C-F bond at Site 5 was the dominant transformation in terms of the relative intensity under irradiation at 280 nm UV-LED. Of note, intermediate B was also observed in the reported study using a low-pressure mercury lamp with UV irradiation at 254 nm [14].

# 3.4. Elucidation of the structure of preliminary intermediates in UV-LED/ $H_2O_2$ treatment

The addition of  $H_2O_2$  may induce a more complicated degradation of CIP, which may be ascribed to the nonselective oxidation of OH'. However, identification of all these intermediates would be time-consuming and somewhat inadvisable. Therefore, a compact two-step procedure was used for the identification of the intermediates in the UV-LED/H<sub>2</sub>O<sub>2</sub> treatment. The first step aimed at discovering the preliminary intermediates, that is, those generated by a one-step transformation from CIP, and the second step focused on seeking the further-transformed intermediates. In the first step, six intermediates, three were also observed in the UV-LED-only experiment, including intermediates A, B and C. However, other three intermediates,  $C_{16}H_{18}FN_3O_2$  (*m*/*z* 304.1),  $C_{15}H_{14}$ -FN<sub>3</sub>O<sub>5</sub> (*m*/*z* 336.1) and  $C_{17}H_{16}FN_3O_4$  (*m*/*z* 346.1), were observed only in the UV-LED/H<sub>2</sub>O<sub>2</sub> experiment (Scheme 2).

Table 4	
Structure elucidation	of stable intermediates.

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No.	Name	Molecular formula	Nominal mass (Da)	Calculated mass (Da) <sup>a</sup>	Mass error range (×10 <sup>-3</sup> Da) <sup>b</sup>	Average isotope ratio difference (%) <sup>c</sup>	MaS (%) <sup>d</sup>	Observed reaction <sup>e</sup>
	CIP	C17H18FN3O3	331.1332	332.1405	2.8-4.6	3.1	97.2	-
1	Α	C <sub>15</sub> H <sub>16</sub> FN <sub>3</sub> O <sub>3</sub>	305.1176	306.1248	-0.5-2.3	5.6	96.3	UV-LED-only and UV/H <sub>2</sub> O <sub>2</sub>
2	В	$C_{17}H_{19}N_3O_4$	329.1376	330.1448	-0.8-1	4.7	96.3	UV-LED-only and UV/H <sub>2</sub> O <sub>2</sub>
3	С	C17H18FN3O4	347.1281	348.1354	1.6-3.7	2.8	81.1A	UV-LED-only and UV/H <sub>2</sub> O <sub>2</sub>
4	D	C <sub>16</sub> H <sub>18</sub> FN <sub>3</sub> O <sub>2</sub>	303.1383	304.1456	-0.9-1.2	3.6	88.1A	UV/H <sub>2</sub> O <sub>2</sub>
5	E	$C_{15}H_{14}FN_{3}O_{5}$	335.0918	336.0990	0.2-2.6	2.7	67.8A	UV/H <sub>2</sub> O <sub>2</sub>
6	F	C17H16FN3O4	345.1125	346.1198	0.6-0.9	5.1	94.3A	UV/H <sub>2</sub> O <sub>2</sub>
7	G	C17H18FN3O5	363.1230	364.1310	-1.7-3.6	4.3	96.1	UV/H <sub>2</sub> O <sub>2</sub>
8	Н	$C_{15}H_{18}FN_3O_3$	307.1332	308.1405	-1.2-2.3	3.6	87.6	UV/H <sub>2</sub> O <sub>2</sub>

<sup>a</sup> Calculated theoretical mass of pseudomolecular ion [M+H]<sup>+</sup> of intermediates.

<sup>b</sup> Difference between measured and calculated theoretical mass.

<sup>c</sup> Difference between measured and calculated isotope distribution.

<sup>d</sup> MaS represents the Matching Score, which indicates the percentage of intensity in regard to MS/MS fragments which can be directly matched with given .mol file (intermediate screening procedure can be seen in Supporting Information).

<sup>e</sup> Intermediates observed in UV and/or UV/H<sub>2</sub>O<sub>2</sub> reactions.



Scheme 1. Proposed generative pathways for the CIP intermediates under 280 nm UV-LED irradiation.

Intermediate  $C_{16}H_{18}FN_3O_2$  (D) had a MW of 303.1383 Da. This intermediate had a 27.9945 Da difference from CIP, and its primary fragments included m/z of 221.0722, 261.1036, 284.1396 and 220.0653 (Fig. S10). In the absence of any carboxyl characteristic fragment (Table S3), it can be concluded that the formation of intermediate D involved the substitution of the terminal carboxyl (-44 Da) by a hydroxyl (+18 Da). The observation of this intermediate inferred that OH<sup>-</sup> can attack the terminal carboxyl of CIP. Intermediate  $C_{15}H_{14}FN_3O_5$  (E) had a MW of 335.0917 Da, which represented a 3.9588 Da difference with that of CIP. The MS<sup>2</sup> spectrum showed that the primary fragments were m/z 316.0928, 298.0820, 275.0541 and 270.0895 (Fig. S11). The fragment m/z316.0928 ( $[C_{15}H_{14}N_3O_5]^+$ ) implied the presence of a C—F bond, and the m/z 298.0820 ( $[C_{15}H_{12}N_3O_4]^+$ ) indicated the presences of both a C—F bond and a hydroxyl. Therefore, it was concluded that the transformation site may be located at the cyclopropyl group or



Fig. 3. Relative intensity and abundance variations of CIP intermediates in 280 nm UV only experiment. The relative intensity indicates the peak area of extract ion chromatogram from MS/MS data, which has a dimensionless unit.



Scheme 2. Proposed generative pathways for the CIP intermediates under 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> treatment.

the piperazine ring. According to the previous study [27], this intermediate was formed by the oxidation of the cyclopropyl group.

Intermediate C<sub>17</sub>H<sub>16</sub>FN<sub>3</sub>O<sub>4</sub> (F) had a MW of 345.1125 Da. Its CID fragments included *m*/*z* 328.1097, 287.0708, 302.1305, 259.0762

(Fig. S12). The fragments m/z 328.1097 and m/z 302.1305 represented the existence of a terminal carboxyl, and m/z 259.0762 showed the cyclopropyl group to be a characteristic fragment. The absence of piperazinyl and fluorine characteristic fragments inferred that the transformation may occur in these two sites.

Eventually, intermediate F was identified to have a substitution of two hydrogen atoms in piperazinyl by an oxygen atom, forming a ketone.

Fig. 4 presents the relative intensity and abundance variations of these six intermediates. Intermediate C had the highest intensity  $(\sim 1.4 \times 10^6 \text{ at } 10 \text{ min})$ , indicating that it may be the dominant product. The observation of these six intermediates implied that all of the peripheral moieties and the quinolone moiety were susceptible to OH<sup>•</sup>. These results were in agreement with those reported in the previous study [25]. However, in the previous study using visible light photocatalysis, the cleavage of the piperazine ring (forming intermediate A) was identified as the dominant transformation mechanism [29]. This difference may be ascribed to the difference in reaction conditions, such as the light sources and the application of catalysts.

#### 3.5. Destruction on the quinolone moiety

Intermediate C was the dominant among the six observed preliminary intermediates. Therefore, the second-step search focused on further products derived from it. After screening, only two further intermediates were identified, including  $C_{15}H_{18}FN_3O_3$  (H, *m/z* 308.1, Fig. S13) and  $C_{17}H_{18}FN_3O_5$  (G, *m/z* 364.1, Fig. S14). Other possible further intermediates may be generated in the degradation process; however, no evidence of these intermediates was observed in the current study.

A consecutive oxidation pathway was proposed (Scheme 3), which included a hydroxyl addition on the C==C (forming intermediate C), cleavage of this C==C and then double oxidation to a ketone (intermediate G), and cleavage and oxidation of the carboxyl (hypothetical product  $C_{16}H_{18}FN_3O_4$ ), ultimately forming intermediate H. Intermediates C, G and H were observed and determined, which provided convincing evidence for this hypothesis.

Fig. 5 presents the relative intensity and abundance variations of these intermediates. The relative intensity of intermediate C rapidly increased to  $\sim 1.4 \times 10^6$  at 10 min and then slowly decreased to  $\sim 0.4 \times 10^6$  within 60 min. The relative intensity of intermediate G increased during the initial 45 min and then slowly decreased. Of note, intermediate H continuously increased for the entire 60 min. These results suggested a generating sequence of intermediates C, G and H, in accordance with the proposed consecutive oxidation pathway. It also implied that UV-LED/H<sub>2</sub>O<sub>2</sub> treatment at 280 nm can destroy the core quinolone structure of CIP.

#### 3.6. Antibacterial activity

To estimate the residual antibacterial activity of a treated CIP solution, a microbial analysis was performed by using *E. coli* as the reference organism. The variation of antibacterial activity was determined by measuring the inhibition halo formed around the micro-drop seeded on the agar plate. Fig. 6 presents the typical results, where the inhibition halo of the UV-LED-only samples slightly decreased as the irradiation time increased. In contrast, the inhibition halo diameter of the UV/H<sub>2</sub>O<sub>2</sub> samples decreased from 27.5 mm (control) to 12.5 mm (30 min), and finally it disappeared after a 45-min reaction, indicating a significant decline in antibacterial activity. The mineralization ratios of these reactions were low; therefore, it can be concluded that the variation of antibacterial activity can be mainly attributed to the modifications of the CIP functional moieties.

CIP is now known to interact with two related but distinct targets, DNA gyrase and topoisomerase IV, inducing further DNA replication failure in bacteria cells. The quinolone moiety is the basic framework of CIP, while the peripheral moieties play additional important roles in antibacterial activity. The carboxyl and carbonyl groups are critical for cleaving or perturbing DNA and



Fig. 4. Relative intensity and abundance variations of preliminary intermediates in 280 nm UV/H<sub>2</sub>O<sub>2</sub> experiment. The relative intensity indicates the peak area of extract ion chromatogram from MS/MS data, which has a dimensionless unit.

# Site 2 Unsaturated double bond in quinolone moiety



Scheme 3. Consecutive oxidation pathway of unsaturated double bond on quinolone moiety and the possible oxidized intermediates under 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> treatment.



Fig. 5. Relative intensity and abundance variations of further intermediates in 280 nm UV/H<sub>2</sub>O<sub>2</sub> experiment. The relative intensity indicates the peak area of extract ion chromatogram from MS/MS data, which has a dimensionless unit.



Fig. 6. Escherichia coli inhibition halo at different times of 280 nm solo-UV-LED irradiation and UV-LED/H<sub>2</sub>O<sub>2</sub> reaction.

are essential for reducing bacteria resistance [30]. The piperazine ring can directly interact with DNA gyrase or topoisomerase IV [31]. The cyclopropyl is part of the enzyme–DNA binding complex and has a hydrophobic interaction with the major groove of DNA [32]. Furthermore, the fluorine at Site 5 can improve the antibacterial activity of CIP. Theoretically, transformations of the carboxyl (intermediate D), the piperazine ring (intermediates A and F), the cyclopropyl (intermediate E) and the C—F bond (intermediate B) all weaken the antibacterial activity of CIP. Of note, because Site 10 at CIP is very close to the site for DNA gyrase (or topoisomerase IV) binding, the addition of a hydroxyl on this site (intermediate C) may also inhibit the antibacterial activity [33].

In the UV-LED-only reaction, intermediate B was the dominant product in regard to the relative intensity (Fig. 3); therefore, the slight reduction of antibacterial activity is mostly ascribed to the formation of intermediate B. Because the C-F bond at Site 5 is not a critical functional group, its elimination had a limited effect on antibacterial activity (Fig. 6e). In contrast, the modifications of all of the CIP functional groups were observed in the UV/H<sub>2</sub>O<sub>2</sub> reaction (Schemes 2 and 3). Intermediate C was the primary product with the highest intensity, and the consecutive oxidation of it to form intermediate H resulted in critical damage to the quinolone moiety, which should show significant inhibition of antibacterial activity. This inhibition was confirmed by the rapid decline of the inhibition halo diameter (Fig. 6i). Of note, there was still 81% of the TOC left after the 45-min UV/H<sub>2</sub>O<sub>2</sub> reaction (Fig. 2c), suggesting that the incomplete degradation of CIP can still induce a significant reduction of antibacterial activity.

### 3.7. EE/O calculation of UV-LED/H<sub>2</sub>O<sub>2</sub> systems

The results of the UV-LED/H<sub>2</sub>O<sub>2</sub> treatment in terms of the removal efficiency and antibacterial activity reduction above were observed to be rather good. However, it is necessary to evaluate its cost. The electrical energy per order (EE/O) values were calculated for the 280 nm UV-LED and UV-LED/H<sub>2</sub>O<sub>2</sub> systems. Detailed information about the calculation method was reported in Ref. [34]. The EE/O value was calculated by taking into account both the electrical energy and chemical oxidant needed in devices [35]. The calculation results are shown in Table S4. The EE/O<sub>-e</sub> values were calculated to be 0.0115 and 0.0027 kWh/m<sup>3</sup>/order, respectively, for the 280 nm UV-LED only treatment and UV-LED/H<sub>2</sub>O<sub>2</sub> treatment. The EE/O value for chemical energy considering the chemical cost was calculated to be 0.0064 kWh/m<sup>3</sup>/order for the 280 nm UV-LED/H<sub>2</sub>O<sub>2</sub> treatment. The EE/O values obtained were significantly lower than the values in experiments using conventional mercury lamps [36].

#### 4. Conclusions

To conclude, different wavelength  $(255-365 \text{ nm}) \text{ UV-LED/H}_2\text{O}_2$  treatments were able to degrade CIP. The UV-LED/H $_2\text{O}_2$  treatment at 280 nm had the highest degradation effectiveness. Three intermediates were identified in the UV-LED-only experiment, and their transformations involved the substitution of the C—F bond by a hydroxyl (intermediate B), the addition of a hydroxyl on cyclohex-

ene (intermediate C) and the breakage of the piperazine ring (intermediate A). In the UV-LED/ $H_2O_2$  experiment, eight intermediates were elucidated, and the transformation involved the oxidation of carboxyl (intermediate D), cyclopropyl (intermediate E) and piperazine (intermediate F). Furthermore, a consecutive oxidation pathway, following the order of CIP and intermediates C, G, E and H, was proposed. Microbial analysis proved that this consecutive destruction of quinolone can reduce the antibacterial activity of CIP. Because quinolone antibiotics all have a similar quinolone moiety, our results implied that 280 nm UV-LED/ $H_2O_2$  can be used as an effective method to degrade quinolones and reduce their antibacterial activity.

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#### Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at http://dx.doi.org/10.1016/j.cej.2016.01.006.

#### References

- [1] Z. Qinqin, J. Ai, W. Yi, L. Hong, W. Kunping, P. Hui, D. Zhaomin, H. Jianying, Occurrences of three classes of antibiotics in a natural river basin: association with antibiotic-resistant *Escherichia coli*, Environ. Sci. Technol. 48 (2014) 14317–14325.
- [2] T.M. Lapara, T.R. Burch, P.J. Mcnamara, D.T. Tan, Y. Mi, J.J. Eichmiller, Tertiarytreated municipal wastewater is a significant point source of antibiotic resistance genes into Duluth-Superior Harbor, Environ. Sci. Technol. 45 (2011) 9543–9549.
- [3] K. Oberlé, M.J. Capdeville, T. Berthe, H. Budzinski, F. Petit, Evidence for a complex relationship between antibiotics and antibiotic-resistant *Escherichia coli*: from medical center patients to a receiving environment, Environ. Sci. Technol. 46 (2012) 1859–1868.
- [4] A. Ludmilla, M. Anastasios, S. Garrison, Inhibition of photosynthesis by a fluoroquinolone antibiotic, Environ. Sci. Technol. 44 (2010) 1444–1450.
- [5] L. Min, D. Wei, H. Zhao, Y. Du, Genotoxicity of quinolones: substituents contribution and transformation products QSAR evaluation using 2D and 3D models, Chemosphere 95 (2014) 220–226.
- [6] Y. Pi, J. Feng, M. Song, J. Sun, Degradation potential of ofloxacin and its resulting transformation products during Fenton oxidation process, Chin. Sci. Bull. 59 (2014) 2618–2624.
- [7] C. Liu, V. Nanaboina, G.V. Korshin, W. Jiang, Spectroscopic study of degradation products of ciprofloxacin, norfloxacin and lomefloxacin formed in ozonated wastewater, Water Res. 46 (2012) 5235–5246.
- [8] M. Mahdi-Ahmed, S. Chiron, Ciprofloxacin oxidation by UV-C activated peroxymonosulfate in wastewater, J. Hazard. Mater. 265 (2014) 41–46.
- [9] X.V. Doorslaer, K. Demeestere, P.M. Heynderickx, M. Caussyn, H.V. Langenhove, F. Devlieghere, V. An, J. Dewulf, Heterogeneous photocatalysis of moxifloxacin: identification of degradation products and determination of residual antibacterial activity, Appl. Catal. B 138 (2013) 333–341.
- [10] L.V. Santos, A.M. Meireles, L.C. Lange, Degradation of antibiotics norfloxacin by Fenton, UV and UV/H<sub>2</sub>O<sub>2</sub>, J. Environ. Manage. 154 (2015) 8–12.
- [11] X.V. Doorslaer, K. Demeestere, P.M. Heynderickx, H.V. Langenhove, J. Dewulf, UV-A and UV-C induced photolytic and photocatalytic degradation of aqueous ciprofloxacin and moxifloxacin: reaction kinetics and role of adsorption, Appl. Catal. B 101 (2011) 540–547.
- [12] J. Di, J. Xia, Y. Ge, H. Li, H. Ji, H. Xu, Q. Zhang, H. Li, M. Li, Novel visible-lightdriven CQDs/Bi<sub>2</sub>WO<sub>6</sub> hybrid materials with enhanced photocatalytic activity toward organic pollutants degradation and mechanism insight, Appl. Catal. B 168 (2015) 51–61.

- [13] P. Huo, Z. Lu, X. Liu, D. Wu, X. Liu, J. Pan, X. Gao, W. Guo, H. Li, Y. Yan, Preparation photocatalyst of selected photodegradation antibiotics by molecular imprinting technology onto TiO<sub>2</sub>/fly-ash cenospheres, Chem. Eng. J. 189–190 (2012) 75–83.
- [14] H.G. Guo, N.Y. Gao, W.H. Chu, L. Li, Y.J. Zhang, J.S. Gu, Y.L. Gu, Photochemical degradation of ciprofloxacin in UV and UV/H<sub>2</sub>O<sub>2</sub> process: kinetics, parameters, and products, Environ. Sci. Pollut. Res. 20 (2013) 3202–3213.
- [15] S. Babić, M. Perišaa, I. Škorićb, Photolytic degradation of norfloxacin, enrofloxacin and ciprofloxacin in various aqueous media, Chemosphere 91 (2013) 1635–1642.
- [16] S. Vilhunen, M. Sillanpää, Recent developments in photochemical and chemical AOPs in water treatment: a mini-review, Rev. Environ. Sci. Bio/ Technol. 9 (2010) 323–330.
- [17] W.K. Jo, R.J. Tayade, New generation energy-efficient light source for photocatalysis: LEDs for environmental applications, Ind. Eng. Chem. Res. 53 (2014) 2073–2084.
- [18] S.H. Vilhunen, M.E. Sillanp, Ultraviolet light emitting diodes and hydrogen peroxide in the photodegradation of aqueous phenol, J. Hazard. Mater. 161 (2009) 1530–1534.
- [19] A. Jamali, R. Vanraes, P. Hanselaer, T.V. Gerven, A batch LED reactor for the photocatalytic degradation of phenol, Chem. Eng. Process. 71 (2013) 43–50.
- [20] C.S. Chiou, C.T. Chang, C.C. Chang, C.Y. Chang, Photodegradation kinetics of formaldehyde using light sources of UVA, composed silver titanium oxide photocatalyst, J. Hazard. Mater. 155 (2008) 164–172.
- [21] R.J. Tayade, T.S. Natarajan, H.C. Bajaj, Photocatalytic degradation of methylene blue dye using ultraviolet light emitting diodes, Ind. Eng. Chem. Res. 48 (2009) 10262–10267.
- [22] E. Repo, S. Rengaraj, S. Pulkka, E. Castangnoli, S. Suihkonen, M. Sopanen, M. Sillanpääa, Photocatalytic degradation of dyes by CdS microspheres under near UV and blue LED radiation, Sep. Purif. Technol. 120 (2013) 206–214.
- [23] M. Izadifard, G. Achari, C.H. Langford, Application of photocatalysts and LED light sources in drinking water treatment, Catalysts 3 (2013) 726–743.
- [24] L. Tang, L. Wang, H. Ou, Q. Li, J. Ye, H. Yin, Correlation among phenyltins molecular properties, degradation and cellular influences on *Bacillus thuringiensis* in the presence of biosurfactant, Biochem. Eng. J. 105 (2016) 71–79.
- [25] A. Taicheng, Y. Hai, S. Weihua, L. Guiying, L. Haiying, W.J. Cooper, Mechanistic considerations for the advanced oxidation treatment of fluoroquinolone pharmaceutical compounds using TiO<sub>2</sub> heterogeneous catalysis, J. Phys. Chem. A 114 (2010) 2569–2575.
- [26] M. Sturini, A. Speltini, F. Maraschi, L. Pretali, A. Profumo, E. Fasani, A. Albini, R. Migliavacca, E. Nucleo, Photodegradation of fluoroquinolones in surface water and antimicrobial activity of the photoproducts, Water Res. 46 (2012) 5575–5582.
- [27] X.X. Zhang, R. Li, M. Jia, S. Wang, Y. Huang, C. Chen, Degradation of ciprofloxacin in aqueous bismuth oxybromide (BiOBr) suspensions under visible light irradiation: a direct hole oxidation pathway, Chem. Eng. J. 274 (2015) 290–297.
- [28] T. An, H. Yang, G. Li, W. Song, W.J. Cooper, X. Nie, Kinetics and mechanism of advanced oxidation processes (AOPs) in degradation of ciprofloxacin in water, Appl. Catal. B 94 (2010) 288–294.
- [29] T. Paul, P.L. Miller, T.J. Strathmann, Visible-light-mediated TiO<sub>2</sub> photocatalysis of fluoroquinolone antibacterial agents, Environ. Sci. Technol. 41 (2007) 4720– 4727.
- [30] G.S. Tillotson, Quinolones: structure-activity relationships and future predictions, J. Med. Microbiol. 44 (1996) 320-324.
- [31] Z. Ma, D.T. Chu, C.S. Cooper, Q. Li, A.K. Fung, S. Wang, L.L. Shen, R.K. Flamm, A. M. Nilius, J.D. Alder, Synthesis and antimicrobial activity of 4H-4-oxoquinolizine derivatives: consequences of structural modification at the C-8 position, J. Med. Chem. 42 (1999) 4202–4213.
- [32] B. Llorente, F. Leclerc, R. Cedergren, Using SAR and QSAR analysis to model the activity and structure of the quinolone–DNA complex, Bioorg. Med. Chem. 4 (1996) 61–71.
- [33] J.M. Domagala, Structure-activity and structure-side-effect relationships for the quinolone antibacterials, J. Antimicrob. Chemother. 33 (1994) 685–706.
- [34] X.X. He, A.A. de la Cruz, D.D. Dionysiou, Destruction of cyanobacterial toxin cylindrospermopsin by hydroxyl radicals and sulfate radicals using UV-254 nm activation of hydrogen peroxide, persulfate and peroxymonosulfate, J. Photochem. Photobiol. A: Chem. 251 (2013) 160–166.
- [35] C.Q. Tan, N.Y. Gao, S.Q. Zhou, Y.L. Xiao, Z.Z. Zhuang, Kinetic study of acetaminophen degradation by UV-based advanced oxidation processes, Chem. Eng. J. 253 (2014) 229–236.
  [36] C.Q. Tan, N.Y. Gao, Y. Deng, Y.J. Zhang, M.H. Sui, J. Deng, S.Q. Zhou, Degradation
- [36] C.Q. Tan, N.Y. Gao, Y. Deng, Y.J. Zhang, M.H. Sui, J. Deng, S.Q. Zhou, Degradation of antipyrine by UV, UV/H<sub>2</sub>O<sub>2</sub> and UV/PS, J. Hazard. Mater. 260 (2013) 1008– 1016.